



## Weed dynamics on Amazonian Dark Earth and adjacent soils of Brazil

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### Abstract

Field trials were conducted on Amazonian Dark Earth soils in the Manaus region, Amazonas, Brazil to assess the composition and impact of weedy vegetation on maize yield. Soil fertility among the Dark Earth varied considerably with differences largely attributable to past-use history. Consequently, maize yield and weed pressure varied among field locations, reflecting these differences in soil fertility in addition to differences in weed reservoirs such as seedbanks. Maize yield in weeded plots was as much as 63 times greater on Dark Earth ( $0.55 \text{ t ha}^{-1}$ ) than on corresponding adjacent soil ( $0 \text{ t ha}^{-1}$ ), and location averages varied from 0 to  $3.15 \text{ t ha}^{-1}$  for Dark Earth. The percentage ground cover of weeds in weedy plots was up to 45 times greater on Dark Earth (65–99%) than on corresponding adjacent soil (2–89%), and species richness was up to 11 times greater on Dark Earth (4–14 species) than corresponding adjacent soil (1–8 species). The relative proportion of annual and leguminous weeds was 32 and 17% greater, respectively, on Dark Earth than adjacent soil, and vegetative sprouting of plants was more common on sites that had been used less intensively in the past. In general, a similar weed community was observed on the different Dark Earth sites, including many species typically associated with environments that have been highly disturbed by human activities, such as *Cyperus* spp., *Phyllanthus niuri*, and *Croton lobatus*. Seedlings from a greater number of species emerged from forested Dark Earth seedbanks (2.1 per flat) than from forested adjacent soil (1.2 per flat). The total number of emerged seedlings was greater for Dark Earth seedbanks (9.1 per flat,  $1,365 \text{ m}^{-2}$ ) than adjacent soil (2.2 per flat,  $330 \text{ m}^{-2}$ ), however the species observed were not likely to be problematic for cropping.

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### 1. Introduction

Amazonian Dark Earth (DE; locally called *Terra Preta de Índio*) patches occur in upland (i.e. non-flooded) environments throughout the Amazon

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Basin, and result from soil formation processes that occurred in pre-Columbian times in areas settled by indigenous people. In general, the significantly higher organic matter and phosphorus (P) contents, increased cation exchange capacity and pH, and lower aluminum (Al) levels (Kern and Kämpf, 1989; Lehmann et al., 2003) allow more intensive agricultural uses of DE soils than adjacent soils (AS). Based on the Brazilian soil classification system, these adjacent soils are considered to be primarily nutrient-impoverished, acidic Latossolos (Salgado Vieira, 1988), which correspond to Oxisols and Ultisols in the USDA soil classification system (Soil Survey Staff, 1997). Where markets can be easily accessed, DE soils can be used for permanent, mechanized crop production requiring high external inputs. In more remote areas, swidden subsistence agriculture is commonly practiced on these soils, with a greater range of crops produced than on AS including maize (*Zea mays* L.), vegetable, and fruit crops. German (2003) reported that farmers consistently rated the performance of a wide variety of crops to be better on DE than AS. This difference in rating was especially striking for maize, with farmers rating its performance at  $2.00 \pm 0.00$  on DE soil and  $0.29 \pm 0.62$  on AS, on a scale of 0 to 2. Fallow periods are reduced by more than 50% on DE compared with AS, and this results in fewer new swidden areas cleared from mature forest on DE than AS (German, 2003).

In 2003, swidden plots planted to maize on both DE and AS were used to determine the richness and diversity of the weed flora, as well as maize and weed biomass production on these two contrasting soils. Working hypotheses included that weed and maize biomass would be greater on DE, and that a different, more diverse weed community would be found above and belowground on these soils compared with AS. Surveys were carried out on 16 DE farms to obtain additional data on weed diversity. The seedbanks of DE and AS soils under forest were sampled to assess the initial weed pressure when swidden agriculture is initiated. Differences in seedbank size between AS and DE were not expected, because sampling both soil types under a forest environment would mask any effects of indigenous disturbance and management incurred centuries ago.

## 2. Materials and methods

Field trials were conducted during the first half of 2003. Maize plantings were established at four locations near Manaus, Amazonas, Brazil (Fig. 1). Each location consisted of one planting on DE and another on AS, for a total of eight plantings. All planting sites were under secondary forest regrowth and not cropped when first visited in August 2002. Paired plantings at a given location were established approximately 200 m apart from each other except at RP, where this distance was 5 km. In the absence of soil analysis data, the location of experimental plots was based largely on soil color; as dark as possible and containing pot shards for DE, and as red or yellow as possible for AS. After slash-and-burn, maize was seeded by hand in all fields between 30 January and 24 February, 2003. The open-pollinated maize cultivar “Solimões” BR 5110, obtained from EMBRAPA (Empresa Brasileira de Pesquisa Agropecuária) was used. Weeds present at the time of seeding were removed by hand or with a hoe. Maize rows were spaced 75 cm apart with 20 cm separating individual plants within the row, to give a density of 66,500 plants  $\text{ha}^{-1}$ . The size of each planting was approximately 12 by 13 m. A randomized complete block design was applied, with three blocks arranged perpendicular to the gradient judged most important at each site, for example slope, distance from road or forest edge. The three treatments evaluated were: (i) weeds only, where no maize was seeded, (ii) weedy maize, where weeds were allowed to compete with the maize until harvest, and (iii) weeded maize, where weeds were removed from plots monthly with a hoe. At each planting site, nine 1.5 m<sup>2</sup> permanent quadrats were established and considered sampling units. Each quadrat spanned two rows of maize with 1.5 m perpendicular to the rows and 1 m along the rows, and included 10 maize plants at seeding. The treatments were applied to an area of approximately 2.25 m × 1.6 m, with the permanent quadrat in the center of this area, and an unweeded row of maize served as a buffer between all treated areas.

After seeding, each site was visited monthly for 4 months to collect crop and weed data. At each sampling, all weed individuals in the 1.5 m<sup>2</sup> permanent quadrats were identified and counted, and the general physiological stage and percentage ground

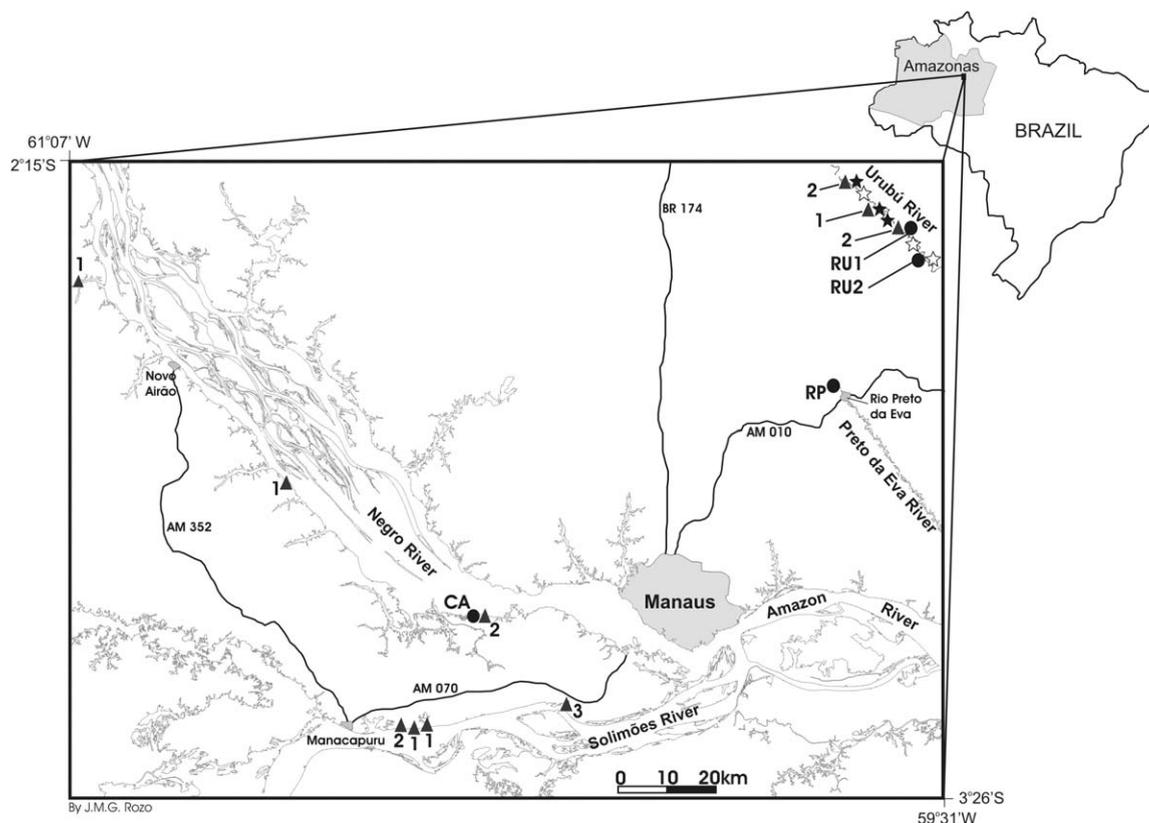


Fig. 1. Location of experimental sites near Manaus, Amazonas, Brazil. Maize trails were conducted at locations CA, RP, RU1 and RU2, and weeds surveys on plots at the locations indicated by (▲). Numbers indicate the number of plots surveyed at each location. Seedbank samples were taken at locations denoted by (★, DE) and (☆, AS).

cover of each species recorded. Total ground cover of weeds was calculated as the sum of percentages for individual species. The number of live maize plants was determined in each quadrat. At harvest, maize plants were cut at soil level, separated into ears with husk and stalks, and fresh weights were recorded. Weeds were separated into the two dominant species, seedlings were separated from sprouts, and the fresh weight of all groupings recorded. When possible, all plant material was brought back to the laboratory, dried at 60 °C for at least 24 h, and weighed.

Weeds were identified with the collaboration of INPA (Instituto Nacional de Pesquisas da Amazônia) and according to Ribeiro et al. (1999) and Lorenzi (2000).

Soil sampling to a depth of 20 cm was carried out using a Dutch auger. Eight samples were collected at seeding from each 12 m × 13 m site, and combined

into one composite sample per site and sampling time. Potassium (K), calcium (Ca), magnesium (Mg) and P were extracted using the Mehlich 1 solution ( $\text{HCl } 0.05 \text{ mol l}^{-1} + \text{H}_2\text{SO}_4 \text{ } 0.0125 \text{ mol l}^{-1}$ ), with available P determined colorimetrically using the ammonium molybdate with ascorbic acid method, K determined by flame emission photometry, and Ca and Mg determined by atomic absorption spectrophotometry. Exchangeable aluminum was extracted with  $1 \text{ mol l}^{-1}$  KCl and the filtrate titrated with  $0.025 \text{ mol l}^{-1}$  NaOH. Total carbon and nitrogen were determined by dry combustion with an automatic Leco CN-2000 CN analyzer (Leco Corporation, St. Joseph, MI, USA). Particle size analysis was carried out using the hydrometer method (Bouyoucos, 1927).

Transects were established on 16 cropped plots on ten farms (Fig. 1), and weeds identified in ten 1.5 m<sup>2</sup> quadrats, separated by 1 m along the transect. Thus, a

total of 15 m<sup>2</sup> were sampled on each of the 16 plots. The plots surveyed were planted to a number of crops, which were at different stages of growth at the time of sampling. A variety of crop management and weed control strategies were used on the different plots. The last weeding prior to the survey ranged from 2 weeks to 2 months. The quadrats sampled for the four DE sites in the maize field experiment, with a total of 13.5 m<sup>2</sup> surveyed per site, were also included in the survey.

For the seedbank study, soil samples were collected on 16–17 April 2003 along the Urubú River, Amazonas, Brazil (Fig. 1), in the latter part of the rainy season that typically lasts from December to May. All sampling locations were under old forest, and based on knowledge from local inhabitants had not been disturbed. For AS sites, this likely meant that these areas were under “virgin” forest, whereas the DE sites had probably not been cropped since abandonment by Indigenous people when their populations were decimated between 1600 and 1700 A.D. Six locations were sampled: Z, A, and S on DE and D, I, and L on AS. At each site, a transect was established approximately 10 m into the forest and parallel to the edge of fallow or cropped fields. At one meter intervals for a distance of 10 m, sampling was carried out along a 1 m line perpendicular to the transect. Using a Dutch auger, four equidistant soil cores were taken along this line on either side of the transect, to a depth of 15 cm. Coarse debris and loose surface roots were removed before sampling. The four samples collected for each of the 10 m locations were combined to yield 10 composite samples per sampling location. The sampling area of 667 cm<sup>2</sup> is within the range used in other similar studies, such as Uhl and Clark, 1983 and Benoit et al., 1989.

Samples were air-dried for one week in the dark before processing. After breaking large soil aggregates by hand, 1 l sub-samples of soil were placed in plastic germination flats. Sixty flats, 10 for each of the six locations, were placed in a greenhouse receiving natural light with flats arranged in a completely randomized design. The photoperiod during the trial was approximately 12 h, and previous data show that temperatures inside the greenhouse during the study period averaged 31 °C but could have reached a daily maximum of 41 °C. Twenty control flats containing 1 l of DE soil of unknown origin autoclaved at 101 kPa

and 120 °C for 30 min were also included. Species emerging in the control flats were considered to have been carried into the greenhouse, which was not fully screened from the outside, and were excluded from the analysis. Flats were watered as required, but not fertilized. Emergence was assessed twice weekly for approximately 59 days. Emerged seedlings were removed and transplanted into separate pots until they could be identified. Due to high temperatures in the greenhouse, shallow soil depths in the flats and logistical limitations, excessive soil drying occurred on several occasions, resulting in 37% of observed seedlings dying before they could be identified. Hereafter, these seedlings are categorized as being “unaccounted”.

Analyses of variance were performed using the GLM procedure in SAS (SAS Institute, 2001), with all terms treated as fixed variables except location in the seedbank study, which was treated as a random variable nested within soil type. After examining residual plots, weed percentage ground cover data were arcsine-square root transformed, counts of weed individuals as well as seedbank emergence counts were square root transformed, and biomass data were log transformed to stabilize variance. In analyses, ramets of species reproducing by rhizomes or root sprouts were considered to be separate individuals. Sorensen’s coefficient of similarity (S) was used to compare weed communities at different sites, according to the equation

$$S = \frac{2c}{s_1 + s_2}$$

where  $c$  is the number of species common to both communities,  $s_1$  is the number of species in the first community and  $s_2$  is the number of species in the second community. Pearson Moment Correlation coefficients were calculated for the various weed and maize parameters.

### 3. Results and discussion

#### 3.1. Soil fertility

DE soils at CA and RU1 were more fertile than their corresponding AS areas, as evidenced by higher pH, P content and base saturation, and lower Al

Table 1

Key soil properties of samples taken in February 2003 after slash-and-burn of secondary growth on four Amazonian Dark Earth (DE) and four adjacent soils (AS) in the central Brazilian Amazon

Parameter	Unit	CA-DE	CA-AS	RP-DE	RP-AS	RU1-DE	RU1-AS	RU2-DE	RU2-AS
pH	(CaCl <sub>2</sub> )	5.02	4.07	4.71	5.03	4.62	4.48	4.51	4.47
P	(mg kg <sup>-1</sup> )	200.1	5.1	44.8	2.8	15.0	3.4	7.5	4.0
BS <sup>a</sup>	(%)	65	2	12	24	20	2	2	4
ALS <sup>b</sup>	(%)	0	19	14	12	10	19	17	17

<sup>a</sup> Base saturation.

<sup>b</sup> Aluminum saturation.

saturation (Table 1). DE soils at RP and RU2, however, exhibited high Al saturation, low P content in the case of RU2-DE, and lower base saturation than the corresponding AS areas (Table 1). Past management may have negatively affected fertility levels, resulting in a fertility gradient among DE sites.

### 3.2. History of study sites

Important differences in past-use history characterized the field sites used in this study. For this reason, the field sites were divided into two groups, based on the intensity of past-use history. The “intensive-use” sites include locations CA and RP (Fig. 1, Table 2), which are closer to markets in Manaus and Rio Preto da Eva, in the case of RP, and where most DE land is under permanent crop production. At the CA-DE site, land use prior to fallowing was intensive as remnants of a road and a suspected dwelling were noticed on the experimental plot. Similarly, the RP-DE site was used as a cattle pasture from 1988 to 1996. The intensive past use of these sites contrasts with the “light-use” sites and may have influenced weed-crop dynamics.

The “light-use” sites include RU1 and RU2. These sites are further away from the city of Manaus and, therefore, market access is substantially reduced. Although some horticultural crops are produced for market on a small scale, subsistence agriculture is the dominant practice on these sites. Experimental plots selected at these “light-use” sites reportedly had been under fallow for shorter periods than plots in the more intensive-use sites, although cropping had not been as intensive.

### 3.3. Maize performance

The effects of soil type, DE or AS, site, and their interaction on maize yield were all highly significant

( $P < 0.0001$ ). While yields were low, maize dry ear biomass was greatest on the more fertile DE sites CA and RU1 (1.4 and 3.2 t ha<sup>-1</sup>, respectively). Lower than expected yields at the CA-DE versus the RU1-DE site were likely due to soil compaction from a road crossing the CA-DE site in the past. At the CA-AS and RU2-AS sites no ears were produced. The CA-DE site, although subjected to high disturbance and soil compaction in the past, did not experience large soil fertility declines as might be expected.

Weeding did not have a significant effect ( $P > 0.05$ ) on overall maize performance. In general, maize did not respond to weed removal, however at the CA-DE site both ear and stalk biomass were, on average, more than double in weeded plots (1.9 t ha<sup>-1</sup> ear) versus non-weeded plots (0.9 t ha<sup>-1</sup> 1 ear). In contrast to other sites, weed pressure at the CA-DE site occurred early but decreased rapidly following the removal of weeds 1 month after seeding when maize was at the six-leaf stage. Yield losses in maize ranging from 2 to 28% have been reported in temperate regions, when weed removal was delayed until four weeks after seeding (Wilson and Westra, 1991; Hall et al., 1992; Hellwig et al., 2002). Higher soil fertility may increase the period during which maize crops can tolerate weeds such as in Nebraska, where the initiation of the critical period for weed control in a maize field was delayed until the six-leaf stage with a nitrogen (N) application of 120 kg ha<sup>-1</sup>, and resulted in an acceptable yield loss of 5% (Evans et al., 2003). When no fertilizer was added, the critical period for weed control occurred at the two-leaf stage of maize. Similarly, maize was better able to tolerate weed pressure with a 220–250 kg N ha<sup>-1</sup> application as opposed to a 100–130 kg N ha<sup>-1</sup> application in field trials in southwestern Ontario, Canada (Tollenaar et al., 1994).

Table 2

Field site description, method of land preparation, and intensity of burn for four Amazonian Dark Earth (DE) and four adjacent soil (AS) sites in the central Brazilian Amazon

Location name	Location code	Standing vegetation <sup>a</sup>	Method of land preparation	Intensity of burn <sup>b</sup>
Costa do Açutuba	CA-DE	More than 10 year old regrowth, dominated by <i>Senna</i> sp. trees 15 cm diameter at base. Suspected location of a dwelling.	Slashed wk of Jan. 21, 2003	Not burned
	CA-AS	Virgin forest, trees 25 m high	Slashed in Feb. 2002, burned in July 2002, hoed before planting	Intense
Rio Preto da Eva	RP-DE	7-Year old regrowth after pasture abandonment, trees sparsely distributed, 5 m high	Slashed wk of Jan. 28, 2003, burned wk of Feb. 10, 2003	Light
	RP-AS	Virgin forest cleared in 1995 and planted to rice, then manioc and fruit trees planted in 1996, site somewhat abandoned at time experiment established	Slashed, and burned and cleaned with a hoe wk of Jan. 10, 2003	Light
Rio Urubu 1	RU1-DE	6-Year old regrowth dominated by <i>Ocotea</i> sp., 10 m high	Understory slashed wk of Jan. 28, 2003, trees cut Feb. 4, burned wk of Feb. 5, branches removed by hand before lightly re-burning the leaf litter and planting	Moderate
	RU1-AS	16-Year old regrowth dominated by <i>Vismia</i> sp., 10 m high	Cleared Sept. 2002, burned wk of Feb. 5, branches removed by hand before lightly re-burning the leaf litter and planting	Moderate
Rio Urubu 2	RU2-DE	3-Year old regrowth dominated by <i>Vismia</i> sp., 10 m high, tree diameter between 3 and 10 cm at base	Cleared and burned Nov. 2002	Moderate
	RU2-AS	8-Year old regrowth slashed in 2001, not previously planted to crops	Re-cut Nov. 2002, burned 7 Feb. 2003, plot cleaned with a hoe	Light

<sup>a</sup> As observed and reported by local inhabitants.<sup>b</sup> Light: tree trunks and most branches left unburned, moderate: most tree branches burned, intense: most plant material burned.

### 3.4. Weed population

Weed populations measured at the different sites varied widely. For example, a weed cover of 100% was obtained on plots at the CA-DE site, while some non-weeded plots at the CA-AS site had a weed cover of 0%. Consequently, there was a significant interaction ( $P < 0.05$ ) between site and soil type for the number of weeds, percentage ground cover and biomass. The percentage ground cover of weeds was strongly correlated to their dry biomass at harvest (Pearson's correlation coefficient = 0.8,  $P < 0.0001$ ), whereas the total number of individual weed plants in the 1.5 m<sup>2</sup> quadrats was poorly correlated with their cover and biomass. Positive weed responses on DE soils were most apparent on the more fertile DE sites CA (98.5% cover) and RU1 (77.3% cover), in comparison to AS at these same sites (2.2 and 26.7%, respectively). Weed cover on AS was greatest at the RP site where *Solanum subinerme* Jacq. was found, which reproduces vigorously by root sprouts. Biomass was generally greater in the plots with only weeds compared to those with weedy maize. As expected, the weeded maize plots had significantly fewer weeds, lower weed ground cover and biomass at harvest than other treatment plots. At harvest, after 4 monthly hand

weedings, weed cover was greater on weeded DE plots at RP (28%) and CA (19%) sites than on similar plots at the other two sites (on average 6.5%) ( $P < 0.05$ ). The seedbank at the CA site was probably very large since within 1 month of the start of the trial, annual weeds and especially *Euphorbia heterophylla* L. were the dominant group and comprised nearly 80% of the ground cover (Fig. 2). At the RP-DE site, however, the dominant species on weeded plots at harvest were members of the Cyperaceae such as *Cyperus* cf. *aggregatus* Endl., *C. cf. diffusus* Vahl, and *Rhynchospora nervosa* (Vahl) Boeck. subsp. *ciliata* T. Koyama, which reproduce primarily vegetatively.

On more fertile DE soils at sites CA and RU1, the better performance of maize reduced weed cover in the weedy maize plots on average by 32 and 58% at CA and RU1, respectively, compared with the plots where only weeds grew, whereas this was not the case at the less fertile DE sites. Early weed emergence at the CA-DE site likely explains why differences in weed coverage between the weed-only and weedy maize plots were not as drastic as for the RU1-DE site, where weed emergence was more gradual. Similar findings were obtained by Jørnsgard et al. (1996) in a cereal field experiment in Denmark. At lower N applications of 30–54 kg N ha<sup>-1</sup>, crop growth and

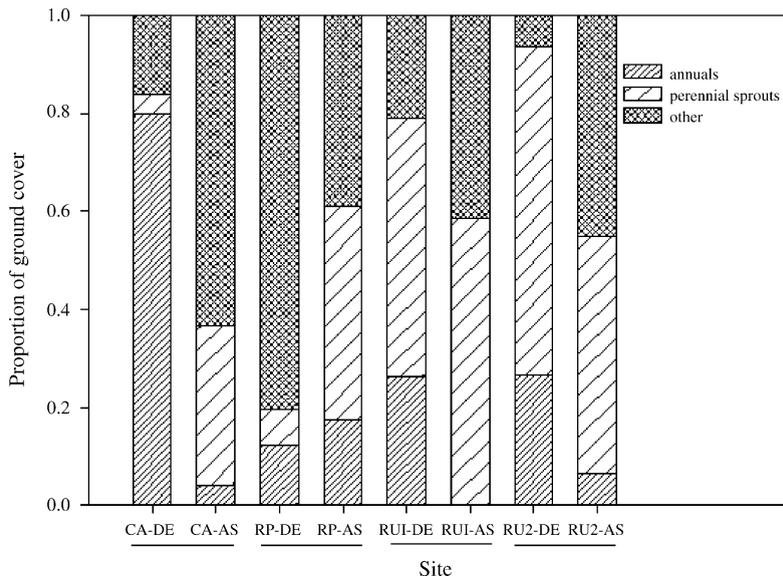


Fig. 2. Proportion of ground cover comprised of annual weeds and perennial sprouts at four Amazonian Dark Earth (DE) and four adjacent soil (AS) sites in the central Brazilian Amazon. The “other” category includes sedges, perennials reproducing by seed, species that can behave as annuals or perennials, and species that could not be identified.

development were reduced and weed growth favored because a greater proportion of light reached the weeds. At higher N application rates of 90–162 kg N ha<sup>-1</sup>, crop growth was improved and plants were able to overtop and shade out the weeds.

The greater weed pressure observed early in the growing season on the CA-DE site was likely due to the high disturbance history of the site and to it not being burned (Table 2). Fire has been demonstrated to reduce seedbanks in proportion to its intensity (Monaco, 1998).

### 3.5. Weed species richness and dominance

A total of 77 plant species were identified at the eight experimental sites. Mean species richness for individual permanent quadrats was greater on DE than AS soils, 8.0 and 4.2 respectively, except at the RU2 site, which had similarly low chemical fertility on both soil types. The greatest number of species was found at the CA-DE site, while the lowest number of species was at the CA-AS site, 29 and 5 respectively.

Species richness has often been observed to follow a unimodal pattern along a productivity gradient, with the highest diversity occurring at moderate productivity levels and lowest diversity found at range extremes (Grime, 1979; Stevens and Carson, 1999). Other factors may also be critical, with disturbance playing an especially important role in facilitating species colonization (Grace, 1999 and references therein). In DE crop production systems, fertility and disturbance are often closely linked. In general, DE soils are located closer to disturbed or open areas such as rivers and dwellings, where plant dispersal and migration are facilitated. In addition, DE soils are subjected to more frequent disturbance, with shorter fallow periods (German, 2003). Weed dispersal and colonization are also assisted by the use of manure and weed seed-contaminated soil amendments, which are most often applied to DE soils and not AS in small-holder agricultural production systems.

Weed communities on DE sites had a more similar species composition ( $P = 0.069$ ) than on AS sites. Many weed species present on sites CA and RP characterized as having an “intensive-use” history were also observed in urban areas of Manaus including roadsides and abandoned lots. Examples of these “cosmopolitan species” include *Euphorbia hirta* L.,

*E. heterophylla*, *Phyllanthus niuri* L., *Cyperus* cf. *aggregatus*, *C.* cf. *diffusus*, and *Microtea debilis* Sw., all of which are native to the Americas, and *Acalypha arvensis* Poepp. and Endl., whose origin is unknown.

On some sites, a single species represented more than 60% of the weed coverage at harvest. Examples include *Spermacoce latifolia* Aubl. at the RU1-DE site and *P. niuri* at the CA-DE site. However, no single weed species was dominant across all field sites. Sixty-two percent or 48 out of 77 weed species were found on only one site, including *E. heterophylla* L., *Sida rhombifolia* L., and *S. subinerme* Jacq. which were each dominant on the site they colonized. Interestingly, *S. rhombifolia* has been reported to be an indicator species of high soil fertility in Costa Rica (de la Cruz, 1994) and we found this species only on the CA-DE site, the most fertile of the sites. Dominance of grasses and sedges was reported on intensively managed soils of Nigeria and Uganda where fertility declines had occurred (Ugen and Wortman, 2001; Akobundu and Ekeleme, 2002). In this study, sedges and grasses were each dominant with >40% of total weed cover on four sites: RP-DE, which was previously used for pasture, and three AS sites: RP-AS, RU1-AS, and RU2-AS. The leguminous woody vine, *Derris amazonica* Killip, was present on all sites except CA, although sometimes at very low densities.

### 3.6. Leguminous weeds

Although differences varied significantly between sites, the number of leguminous individuals and their percentage ground cover was greater on DE than AS (average 21 and 4%, respectively,  $P < 0.0001$ ). Legumes occupied the greatest proportion of ground cover on “light-use” sites RU1 and RU2 where the woody vine *D. amazonica* Killip was dominant. This species relies largely on vegetative sprouting for propagation and, for this reason, is expected at higher densities where burns are less frequent and intense (de Rouw, 1993) and where weeding is less intensive (Kellman, 1980).

High availability of P, Ca, and micronutrients as well as low N availability stimulate the biological fixation of atmospheric N<sub>2</sub>, which gives legumes a competitive advantage on DE (Lehmann et al., 2003) explaining the greater proportion of legume weeds observed on DE than AS. In fact, N availability has

been found to be reduced in DE soils because of high C:N ratios.

### 3.7. Weed reproductive strategy

Annual species constituted a greater proportion of weeds and vegetative cover on DE soils than AS ( $P < 0.05$ ), although differences varied significantly between sites (Fig. 2). This finding is consistent with work by Tilman (1987) and Dieleman et al. (2000) in temperate regions showing that annuals are most abundant on more fertile soils. The greatest differences in the proportion of ground cover comprised by annuals occurred at locations CA and RU1, where fertility differences between DE and AS were greatest.

Weeds entered the reproductive phase earlier on DE than AS sites. During the experiment, an average of 50% of the species occurring on DE sites and 31% of those occurring on AS sites flowered. These differences could be due to the contrasting phenologies of the species observed on each soil type rather than effects of the soil *per se*. More annual species were found on DE soils than AS, and these are expected to flower and set fruit earlier than most perennial species.

Interestingly, weed species found on both soil types did not differ in their timing of flowering and fruit set.

The proportion of biomass from woody sprouting vegetation was greater at sites having “light-use” histories, and did not vary significantly with soil type or weed treatment. Sprouting, as a regenerative strategy, is facilitated on sites with low weeding intensity, less frequent or lighter burns (Kellman, 1980; Uhl et al., 1982; de Rouw, 1993). The substantially lower woody sprout biomass at the RP-DE site was likely due to the site being used for pasture for 8 years, while at CA-DE the presence of a dwelling in the past could explain the lower sprout biomass obtained (Fig. 2).

### 3.8. Weed survey

Ninety-eight species were identified in the weed survey of 20 DE plots including those surveyed for the maize trial. A wider diversity of weeds was recorded, compared to that found in the controlled maize experiment, with 44 of the species found on only one surveyed plot. The most frequent species were *A. arvensis* Poepp. and Endl. (frequency ( $f$ ) = 0.65), *P.*

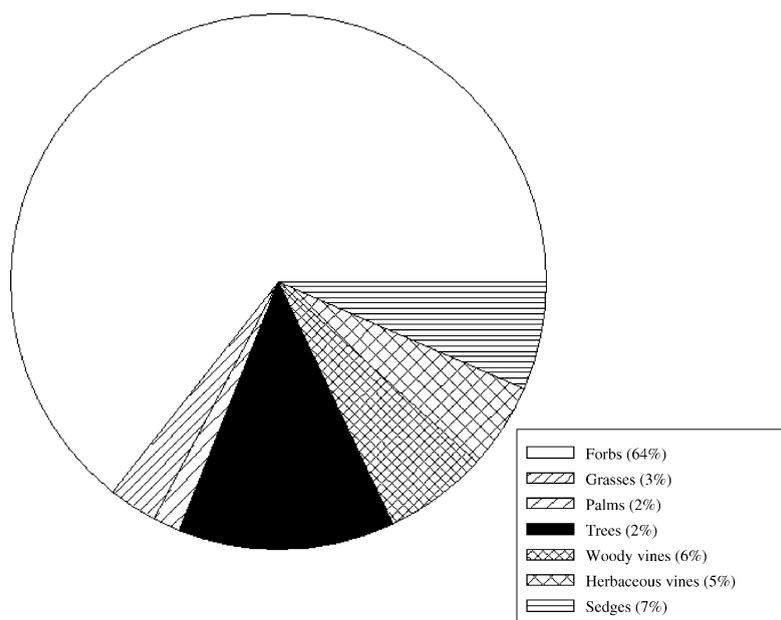


Fig. 3. Frequency ratings by life form for weeds surveyed on 16 cropped Amazonian Dark Earth (DE) plots in the central Brazilian Amazon. Shrubs and epiphytes are not included since each had a frequency of 0.6%.

*niuri* L. ( $f = 0.60$ ), and *Mollugo verticillata* L. ( $f = 0.55$ ). All three may be considered “cosmopolitan” species, as they were frequently observed growing in highly disturbed, urban areas. Thirteen species were in the Euphorbiaceae family and accounted for 24% of frequency recordings. Forbs were the most frequently recorded life form, followed by trees (Fig. 3). Overall, 57% of species surveyed were perennials, 24% were annuals and 19% not known.

### 3.9. Seedbanks

Seedlings from a total of 15 species were recorded: one member each of the Cyperaceae and Leguminosae, three members of *Cecropia*, one member each of *Vismia* and *Fittonia*, *Solanum appressum* K.E. Roe, *Trema micrantha* Blume, and six species groups that could not be identified by name. In general, more seedlings emerged in DE soils than AS (9.1 versus 2.2 per flat, respectively,  $P < 0.05$ ), but there were significant differences ( $P < 0.05$ ) between sampling locations. Similarly, seedlings of more species emerged from DE soils than AS (2.1 versus 1.2, respectively,  $P < 0.1$ ), with no significant effect of sampling location. These values convert to 1,365 and 330 seedlings  $m^{-2}$  for DE soils and AS, respectively.

The mean seedling density for AS corresponds to densities and associated seedbank sizes of 344–862 seedlings  $m^{-2}$  reported by Guevara and Gomez-Pompa (1972) for tropical forest soils. The seedling density for DE soils and its associated seedbank size is within the 2,300 seedlings  $m^{-2}$  range reported by Garcia (1995) for agricultural soils in the tropics, even though the sample area was not under agricultural production. The rate of seedling emergence was greatest between days 21 and 31 and was generally consistent across locations. Emergence rates decreased gradually thereafter (Fig. 4).

The observed differences between DE and AS are likely due to different levels of disturbance associated with the areas adjacent to sampling locations. In a slash-and-burn agricultural system where no tillage is practiced, increased disturbance leads to increased weed biomass and seedbanks when compared with fallowing (Akobundu et al., 1999; Ekeleme et al., 2004). Disturbance is increased on DE because shorter fallows are used (German, 2003), and these soils are often cropped more intensively, at least when markets are accessible. Furthermore, while swiddens on AS are sometimes cleared far from permanent living areas or even on land not directly adjacent to them, dwellings are usually located close to swidden plots on DE, and therefore disturbance is increased. In fact, DE location

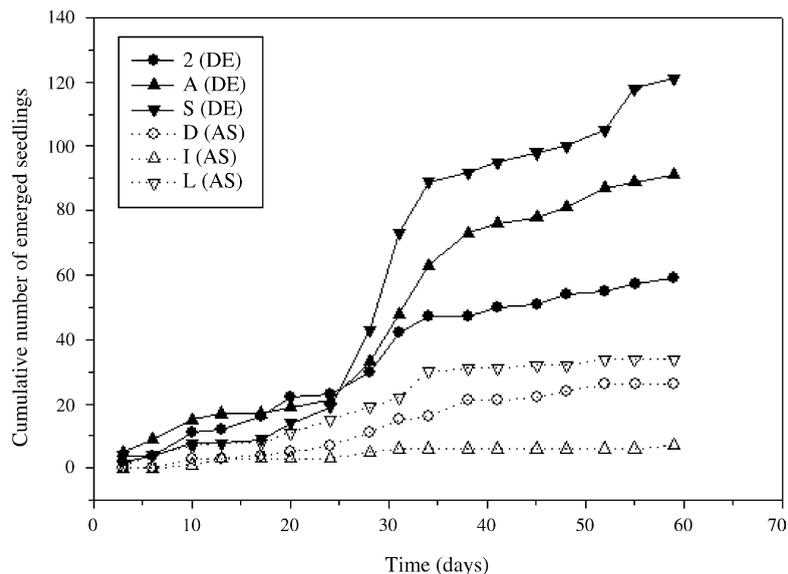


Fig. 4. Cumulative number of emerged seedlings over an 8-week period in a screenhouse for 667  $cm^2$  soil samples from undisturbed soils collected at three Amazonian Dark Earth (DE) and three adjacent soil (AS) locations in the central Brazilian Amazon.

S was located closest to the river (60 m), and had the highest seedling emergence during this trial. A larger population of weeds and early successional species is likely responsible for increased seedbanks in DE sampling areas, via wind and animal dispersal, for example.

A *Cecropia* sp. was the only species found in all locations. Seeds of this early successional tree genus can remain viable in soil for more than 5 years (Holthuijzen and Boerboom, 1982), but manual weeding likely easily controls plants since seedlings cut at soil level would probably not recover. The species identified in this study were not representative of weeds typically occurring in the surveyed DE plots.

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